

Impact of Gypsum Particle Size on Soil Physical Properties of a Saline-Sodic Soil from North Sinai, Egypt

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WATER flow through structured soils can substantially affect local water balance, contaminant transport, and plant-available water. Effect of different gypsum radius on hydraulic conductivity (HC) of saline-sodic soils was assessed. Saline-sodic clay soil from Gelbana Village, Sahl El-Tina plain, Sinai, was mixed with three treatment of gypsum ($\text{CaSO}_4 \cdot \text{H}_2\text{O}$) different in their particle size: T1 "fine" (< 0.5 mm), T2 "medium" (0.5-1 mm) and T3 "coarse" (1-2 mm), and subjected to continuous leaching increments, and these treatments were compared by the control. Soil samples were packed in columns to make a 30-cm height with a bulk density of 1.36 Mg.m^{-3} . Leaching was conducted by ponding with a constant head of 5 cm water. Leaching water was of EC 1.50 dSm^{-1} and SAR of 9.1. Six continuous water increments were performed for each column. To assess and simulate the water flow, to quantify improved management strategies, and to derive updated irrigation standards, the soil-water model HYDRUS-1D code was used. Considerable short-lived increase in HC following addition of gypsum occurred to the soil. It quickly decreased in subsequent leaching increments. The increase amounted to 181, 126 and 117% due to the fine-particles, medium-particles and coarse-particles gypsum, respectively. The increase in HC persisted up to the 4th leachate; and was particularly marked with the fine-particle gypsum, and this increases up to 275% following the 5th leachate compared with the 1st leachate. In soils receiving medium-particle or coarse-particle, increase of the HC was less marked, being up to 56% following 5th leachate increment. The HYDRUS-1D provided reliable simulation results of infiltration rate and cumulative infiltration. Using the model to analyse management options proved an efficient tool for agro-ecosystem assessment.

Keywords: Water flow, Leaching increment, Hydraulic conductivity, HYDRUS-1D code.

Soil salinity is a latent threat in irrigated agriculture countries of the arid and semi arid regions. Almost 400 million hectares of land is worldwide affected by salinization (Bot *et al.*, 2000). Primary salinization contribute with 80% of these lands, while the remaining 20% is "man-made". In addition 1–2% of the irrigated areas, mainly in arid and semi-arid regions, become unsuitable for agriculture yearly (FAO, 2002). Owing to the arid agro-climatic conditions, crop production in Egypt fully relies on irrigation. By the end of the last century,

Egypt irrigated land was estimated at 3 million hectares and the area of salt-affected soils was estimated at 0.8 million hectares are located in the northern, eastern part of the Nile Delta and in Sahl El-Tina plain, Sinai Peninsula. Inadequate management of irrigation has led to considerable salinization of the soils in these regions. Secondary soil salinization by capillary rise of shallow groundwater to the rooting zone plays a major role, entailing yield losses and threatening economic growth and development (Willis *et al.*, 1997, Christen *et al.*, 2001, Singh, 2004 and Murtaza *et al.*, 2006).

Gypsum is the common amendment for saline sodic-soils, because of its low cost, availability and ease of handling. Application of gypsum decreases the ratio of sodicity/salinity in percolating solution and enhances the hydraulic gradient of the soil profile (Miyamoto and Enriquez, 1990, Abdurrahman *et al.*, 2004 and Siyal *et al.*, 2002). Gypsum and phosphogypsum decreased swelling of sodic soils and increased their hydraulic conductivity and infiltration rate and sustained soil structure (Bridge and Tunny, 1973, El-Shanawany, 1985, Ramirez *et al.*, 1999 and Leborn *et al.*, 2002).

Infiltration has long been a focus in agriculture and water research because of its role in land-surface and agricultural irrigation (Milla and Kish, 2006 and Abuhashim, 2011). Large numbers of mathematical models have been developed to evaluate the computation of infiltration. Generally, these infiltration models can be classified into physically based models, semi-empirical and empirical models (Mishra *et al.*, 1999). The semi-empirical and empirical models such as Kostikov and Horton models are often derived from either field and/or laboratory experimental data and are always in the form of simple equations (Lei *et al.*, 1988 and Mishra *et al.*, 2003). However, the semi-empirical and empirical models cannot provide detailed information on infiltration process and their physical impacts. Compared with the semi-empirical and empirical models, the physically-based numerical models can describe the detailed infiltration process and among the physically-based models, is the Richards equation. The Richards equation used the mass conservation law and Darcy's law (Lei *et al.*, 1988). As a physically based model, the Richards equation has been extended into many complex conditions (Brunone *et al.*, 2003, Barontini *et al.*, 2007 and Elmaloglou & Diamantopoulos, 2008). Nevertheless, the Richards equation is strongly non-linear and cannot be solved analytically, especially under complex initial and boundary conditions. Numerical methods such as finite difference and finite element methods have been used to solve the Richards equation (Arampatzis *et al.*, 2001). Based on the finite element method, the HYDRUS-1D model was developed to solve the Richards equation and was used to simulate one-dimensional water movement invariably saturated media (Šimuněk *et al.*, 2005). The soil water model Hydrus-1D was applied to (1) understand and assess the effect of different gypsum radius treatments on the infiltration rate, (2) simulate soil water dynamics under sequential increments of continuous leaching with the

aim to quantify leaching efficiency and (3) simulate improved management strategies in terms of enhancing the hydraulic conductivity.

Material and Methods

Sampling and soil preparation

The saline-sodic soil used in this study was collected from the surface 15-cm layer of a field in Gelbana Village, Sahl El-Tina plain, Sinai. The soil had a history of irrigation with El-Salam Canal (EC 1.5 dSm⁻¹). This region has an extreme continental climate with hot dry summer and rather-wet winter (100-mm of precipitation mainly from November to March). Main crops in this region are wheat, barley, beans and maize, fully dependant on Nile irrigation water of El-Salam Canal. Soil samples were physically and chemically characterized by the standard methods of Black *et al.* (1965) (Table 1). Soil columns were prepared by packing 8.20 kg of sieved (< 2 mm) and air-dried soil into polyvinyl chloride (PVC) cylinders of 40-cm height and 16-cm inside diameter and was set in a vertical position. Soil samples were packed in columns to make a 30-cm height with a bulk density of 1.36 Mg m⁻³. A filter paper disc was put at the bottom of the cylinder with 5-cm layer of acid-washed inert sand (pre-washed with HCl then by distilled water) to form a sand layer in order to prevent removal of soil particles by the flowing water. The top 5-cm on the soil surface gave sufficient space for adding the leaching process. Before packing, the soils were mixed with three treatment of gypsum (CaSO₄.2H₂O) different in there diameter: T1 "fine particles" (< 0.5 mm), T2 "medium paricles" (0.5-1 mm) and T3 "coarse particles" (1-2 mm) and these treatments were compared by a control treatment (without any gypsum supply). The gypsum requirement (GR) was equivalent to 30.6 Mg ha⁻¹ (calculated on a basis of reducing the initial ESP in soil by 90% using the equation given by USDA (1954).

The soil in each column was saturated with water through capillary rise overnight (Baruah, 1997). The infiltration experiment was conducted under ponding condition (*i.e.*, contenuous leaching) with a constant head of 5 cm by a Mariotte bottle. Water used for leaching had an EC of 1.50 dSm⁻¹ and an SAR of 9.1. Six continuous increments were performed for each column (1005 ml per increment) after which the experiment was terminated when the overall total volume of increments was received as leachate. The saturated hydraulic conductivity, K_s is calculated using the Darcy' s equation as follow:

$$\frac{Q}{(A.t)} = -k_s \left(\frac{\Delta h}{L} \right) \quad (1)$$

where, $\left(\frac{\Delta h}{L} \right)$ is the hydraulic gradient, the quantity of water, Q, that flows out of the sample of length, L, and cross-sectional area, A, for a given hydraulic-head, ΔH, is measured for a given time, t.

TABLE 1. Physical and chemical properties of studied soil.

Property	Value
Particle size distribution [%]:	
Clay	44.23
Silt	31.27
Sand	24.50
Texture class	Clay
Soil moisture characteristics [%]:	
Saturation percent	55.23
Field capacity	27.62
Wilting point	13.81
Density [Mgm⁻³]:	
Bulk density	1.36
Total porosity [%]	48.68
Organic matter [g kg⁻¹]	6.20
CaCO ₃ [g kg ⁻¹]	102.3
EC, pH and Soluble ions:	
EC (dSm ⁻¹) [Soil paste extract]	21.38
pH [Soil suspension 1:2.5]	8.44
Soluble ions (mmolc L ⁻¹)	
Na+	175.46
K+	4.66
Ca++	15.22
Mg++	20.16
Cl-	155.20
HCO ₃ -	6.44
SO ₄ =	53.86
SAR	41.72
Exchangeable cations, CEC (cmole kg⁻¹) and ESP	
Na+	12.11
K+	1.35
Ca++	11.93
Mg++	15.22
CEC	40.61
ESP	29.82

*Texture according to the USDA triangle.

Simulation of one-dimension water movement

The HYDRUS-1D code was applied to simulate water movement in variably saturated media and the equation was solved by numerical method (Šimunek *et al.*, 2005). HYDRUS-1D code was based on the one-dimensional Richards equation, The basic water movement equation was described as:

$$\frac{\partial \theta(h,t)}{\partial t} = \frac{\partial}{\partial z} \left[k(h) \left(\frac{\partial h}{\partial z} + 1 \right) \right] \quad (2)$$

where h is the soil water pressure head, θ is the volumetric water content, t is time, z is the vertical coordinate with the origin at the soil surface (negative downward) and $k(h)$ is the unsaturated hydraulic conductivity.

For the studied experiment, the initial condition and upper boundary condition were:

$$h(z,0)=h_i(z) \quad (3)$$

$$h(0,t)=h_0 \quad (4)$$

where $h_i(z)$ is the initial water pressure head through the soil column and h_0 is the soil water potential at soil surface.

The free drainage was considered as the lower boundary condition:

$$\frac{\partial h}{\partial z} = 0 \quad (5)$$

The van Genuchten-Mualem model (van Genuchten, 1980) was applied for the studied experiment:

$$\frac{\theta - \theta_r}{\theta_s - \theta_r} = \left(1 + |\alpha h| \right)^n \quad \text{for } h > 0 \quad (6)$$

$$\theta = \theta_s, h \leq 0 \quad (7)$$

where h is the soil pressure head (cm), θ is the water content ($\text{cm}^3 \text{ cm}^{-3}$), θ^r and θ_s are the residual and saturated water contents ($\text{cm}^3 \text{ cm}^{-3}$), respectively, α , m and n are empirical parameters and $m=1-1/n$.

The HYDRUS-1D code was coupled with the ROSETTA model (Schaap *et al.*, 2001). ROSETTA model implements pedotransfer functions (PTFs) which predict the van Genuchten water retention parameters and the saturated hydraulic conductivity (K_s) in a hierarchical manner from soil textural class information, the soil textural distribution, bulk density and one or two water retention points as input. Within this experiment, the obtained results of the soil textural distribution, bulk density and the measured saturated hydraulic conductivity, K_s , using Darcy's equation under the impact of different gypsum radius treatments and different sequential increments of leaching were implemented in the HYDRUS-1D model to simulate the infiltration rate, cumulative infiltration and the water movement parameters.

Results and Discussion

Analysis of experimental data

Aggregate instability increases the tendency of soils to form structural and depositional crusts (Le Bissonnais, 1996 and Wakindiki & Ben-Hur, 2002), which are common in arid and semi-arid regions. In these regions, crusts are

associated with several desertification factors such as reduced infiltration, enhanced run-off and intensive erosion. The physio-chemical properties of the investigated soil are characterized by a heavy texture with high salinity and sodicity (EC: 21.38 dSm⁻¹; ESP: 29.82 %) (Table 1). The variable hydraulic conductivity (HC) of soil due to sodic conditions depends on the potential of its clays to disperse (Felhender *et al.*, 1974, Abu-Sharar *et al.*, 1987 and Wild, 1993). Shainberg *et al.* (1981) stated that soil dispersion is a function of two closely related variables: salinity and sodicity. At high salt concentration, clay dispersion decreases if the exchangeable sodium percentage (ESP) exceeds 12. Soils with exchangeable Na equal to or greater than 15 % of their exchange capacity and low in soluble salts exhibit poor physical properties, due to Na exerting dispersion of clay colloids and a negative effect on macro aggregate stability (Levy and Torrento, 1995). Saline-sodic soils remain flocculated unless salts are leached from the soil. Dispersion of clay particles and their movement, and possible lodgment in the conducting pores of soils are caused by low levels of electrolytes even with low exchangeable Na (Shainberg *et al.*, 1981, Yousaf *et al.*, 1987 and Lebron *et al.*, 2002).

Figure 1 shows a considerable increase in HC following addition of gypsum. This was short lived and the HC quickly decreased in subsequent leaching increments. Thus hydraulic conductivity was affected by sequential leaching and addition of gypsum fineness (Fig. 1). Gypsum increased HC by 181, 126 and 117% due to the fine-particles, medium-particles and coarse-particles gypsum, respectively (Fig.1). This trend persisted up to the 4th leachate increment and was particularly marked where fine-particle gypsum was applied increases of up to 275% following 5th leachate in comparison with following 1st leachate. In soils receiving medium-particle or coarse-particle gypsum, the increase in HC was less marked, being up to 56% following 5th leachate increment. HC of soils not supplied with gypsum decreased; the decrease was progressive with more leachate increments reaching -51% following the 6th leachate. Such results indicate that gypsum, particularly fine particles, caused an increase in soluble calcium ions, therefore increasing aggregation of soil particles with more Ca⁺² ions on the soil exchange sites. Without gypsum amendment, persistent leaching lead to more deflocculating of clay, thus causing decreased HC.

Simulation of one-dimensional water movement

As the current results reveal that the HC is much higher under the treatment of T1 Gypsum than the other treatments, only the results of the numerical model HYDRUS-1D will be presented here to evaluate the sequential leaching increments under the impact of the fine gypsum (T1) on the water movement. Implementing the results of texture analysis, bulk density, field capacity and the measured saturated hydraulic conductivity for investigated soil in the HYDRUS-1D model, the hydrological parameters (Q_r , Q_s , α , n) can be optimized.

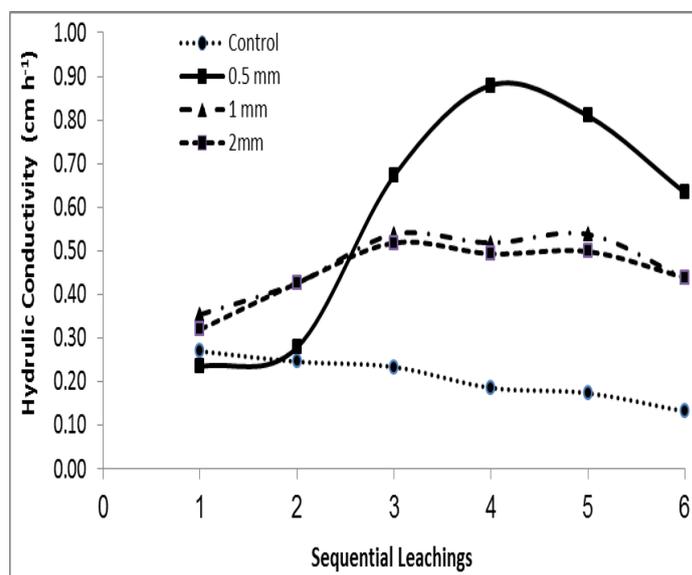


Fig.1. Hydraulic conductivity of the saline sodic (21.38 dSm^{-1} ; 29.82 SAR) clay soil following a surface application of gypsum (30.6 Mg ha^{-1} equivalent) and followed by leaching with water of 1.50 dSm^{-1} and 9.1 SAR .

The results of Fig. 2 show differences in the saturated HC (K_s) values between the leaching increments for soils receiving the finer particle gypsum. The differences are not large, but it supports the general supposition of the water flow as a function of time with depth. The results reveal that simulated distribution of hydraulic conductivity under the 4th leaching increment (K4) was abundant and larger than the other increments. It required 5 hr for the entire top soil surface (0-30 cm) to be fully saturated under the impact of the 4th leaching increment (K4), while it requires 18, 16, 7, 6 and 7 hr for the entire top soil surface (0-30 cm) to be fully saturated under the impact of the other continuous leaching increments K1, K2, K3, K5 and K6, respectively. Such change in K_s may have been due to attributed to internal swelling in the soil which would be reversible and would play a major role in controlling the hydraulic properties of this soil. The changes in swelling and HC could be related to the ionic strength effect. With each subsequent leaching increment, there would be less gypsum in the soil. This suggests a possible pore-plugging by dispersed clay and slaked particles during the pre-leaching periods. Such a decrease in HC could be largely irreversible (Yousaf *et al.*, 1987 and Curtin *et al.*, 1994). Relationship between clay dispersion and K_s can be utilized to provide an index for sodicity hazards and for predicting reduction in hydraulic conductivity.

Fig. 2. Comparison of simulated hydraulic conductivity (K_s) by HYDRUS-1D code with soil depth. The Six continuous increments; K1, K2, K3, K4, K5 and K6 show this simulation for soils receiving the finer particle gypsum.

For the following results, only measurements during the first 8 hr of the experiment will be displayed in the analysis below. Using the numerical model HYDRUS-1D, the impact of the leaching increments on the infiltration rate was studied through assessing the cumulative infiltration curve in soils receiving the fine particle (< 0.5 mm) gypsum. The cumulative infiltration curve under the *Egypt. J. Soil Sci.* **53**, No. 1 (2013)

different leaching increments is shown in Fig. 3. Infiltration with the 4th leaching increment (K4) was much larger than the others. These results reflect the effect of water retention. The total cumulative infiltration after five hours of leaching increment was 1.23, 1.36, 2.33, 2.75, 2.61 and 2.24 cm for K1, K2, K3, K4, K5 and K6, respectively. These results correspond well with the soil water storage capacity (Fig. 4) which reflects the same trend of increasing the water storage under the 4th leaching increment (K4) compared with the other increments.

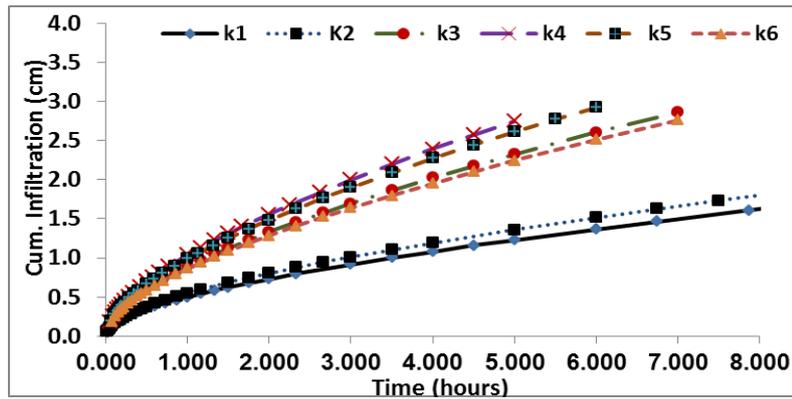


Fig. 3. Simulated cumulative infiltration by the HYDRUS-1D code versus time. The Six continuous increments; K1, K2, K3, K4, K5 and K6 show this simulation for soils receiving the finer particle gypsum.

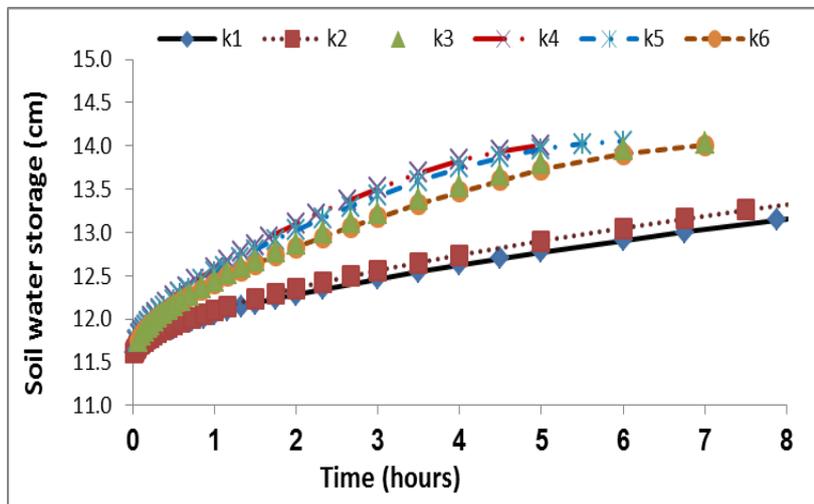


Fig. 4. Comparison of simulated water storage by HYDRUS-1D code versus time. The Six continuous increments; K1, K2, K3, K4, K5 and K6 show this simulation for soils receiving the finer particle gypsum.

The volumetric water content was considerably higher under the 4th leaching increment compared to the other sequential leaching increments in the top soil surface (0-30 cm). Fig. 5 shows water content (v/v) distribution in the top soil surface at 10 cm by using the numerical model of the HYDRUS-1D (Fig. 5). The volumetric water content after five hours of infiltration was constant under different sequential leaching increments (Fig. 5), with a value of 0.453, 0.456, 0.466, 0.469, 0.468 and 0.466 [cm³cm⁻³], for K1, K2, K3, K4, K5 and K6, respectively. The rate of the dissolution of gypsum in water would increase with increase in its surface area (decreasing with gypsum finnes). Kemper *et al.* (1975) observed that the first order dissolution coefficient of gypsum fragments in water at a given flow rate increased more rapidly as the fragment size became smaller (as the surface area increased).

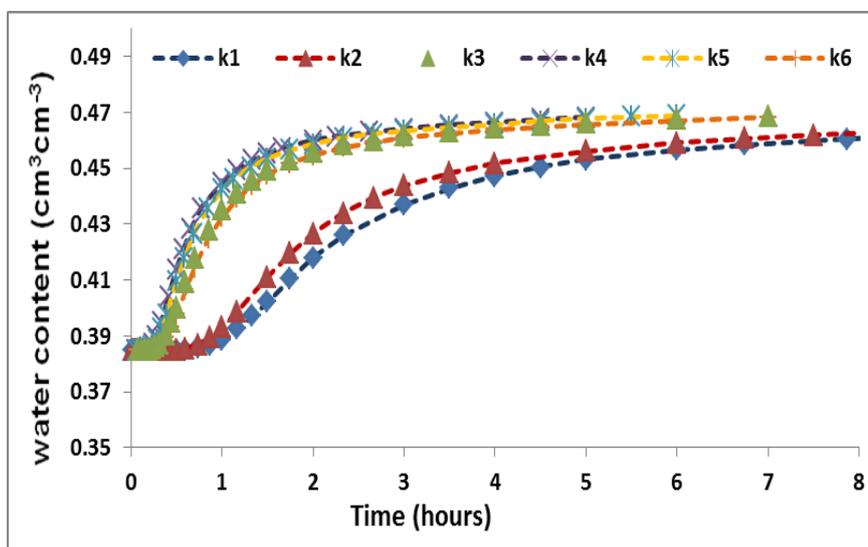


Fig. 5. Calculated water content by HYDRUS-1D code versus time at observation node (10 cm) depth. The Six continuous increments; K1, K2, K3, K4, K5 and K6 show this simulation for soils receiving the finer particle gypsum.

Keisling *et al.* (1978) displayed an exponential relationship between the first order rate coefficient and the surface area to accurately predict dissolution at various flow rates. This suggests that in addition to the differences in the surface area the sources differ in their reactivates (Barton and Wilde, 1971). In solutions containing Ca²⁺ and/or SO₄ ions, the dissolution rate of gypsum was reported by Kemper *et al.* (1975) to decrease due to a Ca²⁺ and SO₄ ion effect. When the HC was less than 1.6 mm h⁻¹, the saturated gypsum solution would be shown in the leachate as long as solid gypsum is present in the soil. In some soils under field conditions, Bolan *et al.* (1991) noted that water often flows preferentially through cracks and macropores and hence a saturated concentration of gypsum

may not be maintained in such soils. The decrease in K_s obtained in the current study could be attributed to clay dispersion and partially plugging of the water conducting pores (Abuhashim *et al.*, 2009 and Lebron *et al.*, 2002).

Conclusion

Dynamics of soil-water-movement in fields could be simulated accurately with HYDRUS-1D model. Using the model to compare, simulate and analysis the management options provide a powerful tool for agro-ecosystem assessment. Although soils of Sahl El-Tina plain are characterized as an area with continuous secondary salinization, the overall soil salinity in this region seems to have so far been kept under control by pre-season salt leaching. With addition gypsum, particularly the fine-particle material; there would be a higher solubility calcium sulfate, *i.e.*, calcium ions, therefore increasing aggregation of soil particles. With removal gypsum from the soil, persistent leaching would be associated with increased presence of Na^+ ions so as to render it dominate the soil exchange complex which would lead to deflocculating, thus causing decreased HC. Therefore, for maintaining a preferable soil structure in saline-sodic soils to help obtaining efficient leaching, gypsum applied in reclamation such soils should be of fine particles.

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تأثير حجم حبيبات الجبس على خواص التربة الفيزيائية لأراضي ملحية صودية من شمال سيناء، مصر

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أقيمت تجربة أعمدة على تربة ملحية صودية مأخوذة من منطقة سهل الطينة - شمال سيناء - مصر، لدراسة تأثير استخدام حبيبات الجبس ذات اقطار مختلفة على التوصيل الهيدروليكي (HC) وخواص التربة الطبيعية (0) تم استخدام ثلاثة أقطار من الجبس الزراعي ($\text{CaSO}_4 \cdot \text{H}_2\text{O}$) كما يلي: القطر الأول أقل من 0.5 ملم (T1) والقطر الثاني من 0.5 إلى 1 ملم (T2) بينما القطر الثالث كان يتراوح بين 1 إلى 2 ملم (T3)، خلطت كميات الجبس المحسوبة من كل قطر خطأ متجانساً مع أعمدة تربة بارتفاع 30سم وقطر 16 سم وكثافة ظاهرية 1.36 جم/سم³ وأجريت عملية الغسيل بامرار 6 جرعات بماء ذو ملوحة 1.5 ملليموز/سم ، SAR 9.1 حيث مُررت المياه تحت ضغط هيدروليكي متغير مقدارة 5 سم⁰ تم ادخال النتائج المتحصل عليها إلى الموديل الرقمي HYDRUS-1 لتقييم ومحاكاة حركة المياه تحت ظروف المعاملات المستخدمة لتحديد استراتيجيات إدارة وري الأراضى الملحية الصودية.

أظهرت النتائج المتحصل عليها حدوث زيادة كبيرة في قيم HC وذلك لمدة قصيرة الأجل بعد إضافة الجبس إلى التربة وانخفضت قيم HC بسرعة بزيادة جرعات الغسيل المتوالية. وبلغت هذه الزيادة إلى 117، 126، 181 % باستخدام الجبس الزراعي ($\text{CaSO}_4 \cdot \text{H}_2\text{O}$) ويمكن ترتيب المعاملات طبقاً لمقدار الزيادة كما يلي: $T1 > T2 > T3$.

استمرت الزيادة في قيم HC حتى الجرعة الرابعة، وكان ذلك ملحوظاً بشكل خاص مع المعاملة T1، وهذه الزيادة تصل إلى 275% بعد جرعة الغسيل الخامسة مقارنة مع جرعة الغسيل الأولى (0) بينما فى التربة التى تتلقى حبيبات الجبس المتوسطة والخشنة الحجم كانت الزيادة فى قيم HC أقل وضوحاً و تصل إلى 56% بعد جرعة الغسيل الخامسة (0) أدى استخدام الموديل الرقمي HYDRUS-1D إلى تحديد وتقييم حركة المياه للاعماق المختلفة لعمود التربة تحت تأثيرات المعاملات المستخدمة ، وكذلك أدى إلى تحديد الخيارات الفعالة لإدارة النظم الزراعية تحت المعاملات المختلفة (0).